

Population cycles caused by selection by density dependent competitive interactions

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Abstract Several animal species have cyclic population dynamics with phase-related cycles in life history traits like body mass, reproductive rate, and pre-reproductive period. Although many mechanisms have been proposed there is no agreement on the cause of these cycles, and no population equation that deduces both the abundance and the life history cycles from basic ecological constraints has been formulated. Here I deduce a population dynamic equation from the selection pressure of density dependent competitive interactions in order to explain the cyclic dynamics in abundance and life history traits. The model can explain cycles by evolutionary changes in the genotype or by plastic responses in the phenotype. It treats the population dynamic growth rate as an initial condition, and its density independent fundament is Fisher's (1930) fundamental theorem of natural selection that predicts a hyper-geometrical increase in abundance. The predicted periods coincide with the cyclic dynamics of Lepidoptera, and the Calder hypothesis, which suggests that the period of population cycles is proportional to the 1/4 power of body mass, follows from first principles of the proposed density dependent ecology.

Keywords: Population dynamics, evolutionary dynamics, interference competition, fluctuations, Lepidoptera.

1 Introduction

The widespread occurrence of cycles in the abundance and life history characters of animal species has remained an unsolved problem in animal ecology (e.g., Boonstra and Krebs 1979; Lidicker and Ostfeld 1991; Stenseth and Ims 1993; Chitty 1996; Krebs 1996). According to classical population dynamic theory, which is based on density regulation and Malthusian growth for populations in density independent environments, population cycles are the exception that require some sort of delayed density dependence. At present at least 22 mechanisms have been proposed to explain delayed density dependence and the cyclic dynamics (Batzli 1996), but no clear conclusions have been made. The proposed hypotheses have been based on mechanisms extrinsic or intrinsic to the population, with the most

widespread extrinsic hypothesis being predator-prey interactions in their broadest definition that include both host-parasitoid and plant-herbivore interactions. The most widespread intrinsic hypotheses are those of over-compensation and density dependent changes in individual quality.

Predator-prey interactions may be the most popular hypothesis for population cycles in small mammals and insects (e.g., Akçakaya 1992; Hanski et al. 1993; Hanski and Korpimäki 1995; Krebs et al. 1995; Norrdahl 1995; Berryman 1996; Jedrzejewski and Jedrzejewski 1996; Turchin et al. 2000). Despite the extensive research effort devoted to this hypothesis Berryman (1996) notes that the best documented cycle in insects lacks a firm predator-prey interaction. And Krebs (1996) concludes that “Changes in the food supply have never been shown to be necessary or by itself sufficient to cause cycles”, and that “there is not a single field experiment to show that predator removal has any impact on cycles of voles or lemmings.” Especially the phase-related cycles in life history traits are not expected from the extrinsic hypotheses unless, of course, they are induced by another mechanism and follow as a consequence and not a cause of cyclic population dynamics (Lidicker and Ostfeld 1991; Oksanen and Lundberg 1995; Oli 1999).

The intrinsic hypothesis of over-compensatory dynamics has been studied theoretically by May and Oster (1976), Sandefur (1990), and Strogatz (1994). This mechanism can explain periods for up to two generations (e.g., Tuljapurkar et al. 1994), and over-compensatory dynamics have been found in semi natural populations as diverse as ungulates (Grenfell et al. 1992) and insects (Hassell et al. 1976; Gurney et al. 1980; Desharnais and Liu 1987; Costantino et al. 1995). But it is generally agreed that over-compensation cannot explain the general tendency for cyclic dynamics with periods of more than two generations (Turchin 1990; Witteman et al. 1990; Turchin and Taylor 1992; Ginzburg and Taneyhill 1994).

It has often been suggested that other intrinsic mechanisms that are based on density dependent changes in

individual quality can explain population cycles with periods of more than two generations. These intrinsic hypotheses include changes caused by physiological (Christian 1950; Boonstra 1994), selectional (Chitty 1960, 1967), behavioural (Charnov and Finnerty 1980), and maternal effect (Wellington 1965) mechanisms. Compared with the extrinsic hypotheses the intrinsic models may seem superior in the way that they tend to incorporate the phase-related changes in life history traits. But comparing the different intrinsic models Stenseth (1981, 1985, 1995) concludes that they generally fail to explain the occurrence of population cycles and/or that they are burdened by unrealistic assumptions.

It has been argued that the selection mechanism is experimentally rejected because the level of heritable variation seems to be so low that evolutionary changes in the genotype cannot cause cyclic population dynamics (Boonstra and Boag 1987; Boonstra and Hochachka 1997). However, the phenotypic response to selection may not only arise from evolutionary changes in the genotype. The phenotype may also change due to inherited environmental effects, where an epigenetic inheritance system transfer a plastic phenotypic response from a parent to an offspring generation (e.g., Jablonka and Lamb 1989, 1998; Rossiter 1996). Maternal effect is one of many possible inherited environmental effects, and the maternal effect hypothesis for population cycles has recently received much attention (e.g., Rossiter 1991, 1992, 1994; Ginzburg and Taneyhill 1994, 1995; Boonstra and Hochachka 1997; Ginzburg 1998; Inchausti and Ginzburg 1998). Compared with earlier versions of the intrinsic hypotheses the maternal effect hypothesis is more promising because it is now formulated into a population equation with dynamics that resemble the cycles of Lepidoptera and small mammals (Ginzburg and Taneyhill 1994, 1995; Ginzburg 1998; Inchausti and Ginzburg 1998). Although it has not been shown empirically whether cyclic and density dependent changes in inherited environmental effects are generating population cycles in natural populations, the models of Ginzburg and Taneyhill (1994, 1995) and Inchausti and Ginzburg (1998) show that it is a plausible hypothesis. Their studies, however, do not suggest from what ecological factors the plastic phenotypic response arises. From a mechanistic point of view it is the density dependent factor that induces selection for between generation changes in life history parameters that is the ultimate factor that may explain cyclic population dynamics by density dependent changes in individual quality. On a longer evolutionary time scale such factors can induce the evolution of an inheritable

phenotypic response, especially, if the density dependent environment fluctuates with a period that exceeds the generation time of the organism (Lachmann and Jablonka 1996). And if the potential for an inheritable phenotypic response has evolved by these means, then, it is the density dependent changes in the same ecological factor that is most likely to induce a plastic between generation response.

To fully understand the potential for cyclic dynamics by density dependent changes in individual quality we need first to identify a density dependent factor that selects for cyclic and density dependent changes in individual quality and population abundance. But despite of several attempts (e.g., Dekker 1975; Stenseth 1978, 1981; Thue Poulsen 1979; Hunt 1982) the original selection hypothesis of Chitty was never formulated into a realistic population dynamic equation (Stenseth 1981, 1985, 1995). In this paper I show that selection by the density dependent competitive interactions among individuals in populations can be the ultimate ecological factor behind cyclic animal dynamics. This is done by a selection model where the across generation response in population parameters to the selection pressure of density dependent competitive interactions can arise in two ways. The first is a plastic response where an epigenetic inheritance system adjusts the phenotype to the selection pressure of competitive interactions. The second is evolution by natural selection as originally proposed by Chitty (1960). Provided a sufficiently high level of additive genetic variation in population parameters it is shown that selection by density dependent competitive interactions can generate population cycles in the absence of a plastic phenotypic response. If instead the level of genetic variation is negligible a plastic response is needed before the cycle may arise.

1.1 Selection by density dependent competitive interactions

The model in this paper is part of a larger theory where the selection pressure of density dependent competitive interactions is used to deduce major life history transitions in mobile organisms (Witting 1995, 1997, 2000). The theory deals with the evolution of life history characters like reproductive rate, metabolic rate, body mass, exponents of body mass allometries, senescence, sex ratios, and sexual and eusocial reproduction. In this paper I examine the population dynamics around an evolutionary equilibrium of this theory. This is done by considering the effects that selection by density dependent competitive interactions have on traits like body mass and the population dynamic growth

rate. The implications of a theoretical deduction of the exponents of ten body mass allometries (Witting 1995, 1998) is also considered in the sense that the Calder hypothesis (Calder 1983, 1984), which states that the period of cyclic dynamics is proportional to the 1/4 power of body mass, is deduced from first principles of the proposed density dependent ecology.

Classical life history models (reviewed by Roff 1992; Stearns 1992; Charnov 1993; Bulmer 1994; Charlesworth 1994; Kozłowski 1999), including models of density dependent selection (reviewed Mueller 1997), were developed independently of competitive interactions and frequency dependent selection. Selection by competitive interactions has instead been described under the concepts of game theory and evolutionary stable strategies (ESSs) (Maynard Smith and Price 1973; Axelrod and Hamilton 1981; Maynard Smith 1982; Vincent and Brown 1988; Vega-Redondo 1996; Dugatkin and Reeve 1998; Hofbauer and Sigmund 1998), with more recent considerations on evolutionary convergence (Eshel 1983; Taylor 1989; Christiansen 1991) and evolutionary branching (Metz et al. 1992, 1996; Dieckmann 1997; Eshel et al. 1997; Geritz et al. 1997, 1998; Kisdi 1999). Like most of this work I deal with the evolution of traits that indicate the competitive ability of the organism.

To understand how selection by density dependent competitive interactions can generate cyclic dynamics consider first selection by the classical life history theory that disregard the effects of competitive interactions. This theory defines fitness as the intrinsic population dynamic growth rate, also known as the intrinsic Malthusian parameter. Dependent upon the density dependent state of the population the classical theory is often understood in terms of either r - or k -selection, where respectively the maximal population dynamic growth rate and/or the carrying capacity increases. More generally, however, as predicted by Fisher's fundamental theorem of natural selection (Fisher 1930; Price 1972; Witting 2000), classical life history selection is given by an absolute increase in the intrinsic growth rate (this increase is also known as a partial increase in the absolute growth rate). Thus, as competitive traits tend to be related to the intrinsic growth rate by an energetic trade-off (Witting 2000), from the classical theory we expect a continuous decline in competitive traits independently of the density dependent ecology. This prediction will not generate cyclic dynamics that require that the direction of the evolutionary changes in the intrinsic growth rate is density dependent.

With density dependent competitive interactions

there are instead two opposing forces of selection on the intrinsic growth rate and the competitive traits. The first is classical r - and k -selection that favours the intrinsic growth rate at the cost of the competitive traits. The second is the selection pressure of density dependent competitive interactions that favours the competitive traits at the cost of the intrinsic growth rate. If the population abundance and, thus, the level of interference competition is sufficiently high the latter force is stronger than the classical force resulting in overall selection for an increase in the competitive traits at the cost of the intrinsic growth rate. If instead the population abundance and the level of interference are sufficiently low the force of r - and k -selection will dominate and the intrinsic growth rate will increase. At the evolutionarily determined population dynamic equilibrium the population abundance is exactly so high that the two selection forces are balanced against one another generating no overall selection. However, if the population is not at equilibrium the density dependent selection will generate across generation changes in the population dynamic growth rate. And dependent upon the magnitude of the population's response to selection this form of delayed density dependence may generate cyclic population dynamics.

It has earlier been shown that selection by competitive interactions may induce evolutionarily driven population dynamic cycles for the case of density independent competitive interactions (Maynard Smith and Brown 1986; Härdling 1999). This is possible when evolution is mutation-limited in the sense that mutations occur so infrequently that a mutant is excluded or spread to equilibrium before a new mutant arises. In this case a large variant may be invaded by extreme small mutants that do not experience the cost of a large size. The small mutant may then spread to fixation so that the size of the organism can increase until a new invasion by a small mutant is possible. These cycles, however, may not be applicable to higher organisms. Invasions by extreme types are generally not possible with sexual reproduction and quantitative inheritance where a unimodal phenotypic distribution tends to be maintained (Matsuda and Abrams 1994). A period of approximately 4000 generations, as predicted by Härdling (1999), is also far beyond the periods usually considered in connection with cyclic population dynamics. The population cycles predicted in the present paper can have much faster periods and they are applicable to higher organisms in the sense that they operate through evolutionary cycles in the mean of unimodal phenotypic distributions.

2 The model

In the absence of selection the density dependent dynamics of a species with non-overlapping generations can be described as

$$\tilde{N}_{t+1} = \tilde{N}_t \lambda_m f(\tilde{N}_t) \quad (1)$$

where λ_m is the maximal per generation growth rate, \tilde{N} is the population abundance that is larger than or equal to one for all extant populations, and $f(\tilde{N})$ is the density regulation function that declines monotonically from one to zero as the abundance increases from one to infinity. Instead of representing density regulation by the general model of eqn 1, let me focus on the process that is linear at the logarithmic scale of population dynamics. The density regulation function can then be defined as

$$f(\tilde{N}) = \begin{cases} \nu \tilde{N}^{-\gamma} & \text{if } \tilde{N} > \dot{\tilde{N}} \\ 1 & \text{if } \tilde{N} < \dot{\tilde{N}} \end{cases} \quad (2)$$

where $\dot{\tilde{N}} = 1$ is the abundance where the effects of density regulation vanish, γ is the parameter that defines the curvature of the density regulation function, and $\nu = \dot{\tilde{N}}^\gamma$. Thus, if the population abundance is scaled as $N = \tilde{N} \nu^{-1/\gamma}$, and if we assume that $\tilde{N} \geq \dot{\tilde{N}}$, which is reasonable as long as we consider only local perturbations of the population equilibrium, we obtain the following population dynamic equation

$$N_{t+1} = N_t \lambda_m N_t^{-\gamma} \quad (3)$$

that forms the basis of my study. Let, for this model, the density regulation parameter γ be defined as $\gamma = \gamma_\alpha + \mu\gamma_\nu$, where γ_α is density regulation by resource exploitation, γ_ν is density dependence in the level of interference competition, and μ is the cost per unit interference.

The model of eqn 3 is based on the assumption that the maximal growth rate (λ_m) is constant. But due to the action of selection we expect that the maximal growth rate is both time and density dependent. To describe the expected changes in λ_m and their implications for population dynamics I partition the phenotype into demographic and competitive traits. Let the demographic traits be lifetime reproduction (R) and the probability that an offspring survives to reproduce (p), which define the population dynamic growth rate $\lambda = pR$. And let the competitive traits be traits that individuals, or variants, can use to dominate other individuals, or variants, during competitive encounters.

A common feature of the competitive traits is that they are connected to the population dynamic growth rate by a trade-off. Let $q = wBT_p$ represent competitive quality as defined by the product between the body mass (w), the metabolic rate per unit body mass (B), and the pre-reproductive period (T_p). The body mass can be a competitive trait because it can be selected so that the individuals with the larger body masses can dominate the smaller individuals during competitive encounters. The same is true for the metabolic rate where the individual with the highest metabolic rate has the potential to allocate the largest amount of energy into the competitive encounter. The pre-reproductive period may also represent competitive quality because it is during this period that the offspring has the time to learn the behaviour that is needed to dominate other individuals. The idea of treating these traits as competitive traits is not to suggest that the traits might not have other functions in natural organisms. It is only to show that the evolution of these traits, and their association with population dynamics, is relatively easily explained by density dependent competitive interactions.

If we assume that the offspring are reared by the parents the competitive quality q may also be interpreted as the energy that the parents invest per offspring. Actually this energy is better described as $w+q$ because q represents only the energy that the offspring metabolises during the pre-reproductive period, while the energy invested in an offspring includes also the energy contained in the body mass of that offspring. But, for the sake of simplicity I assume that q resembles the energy invested per offspring. Thus, we expect that

$$\lambda_m = \rho/q \quad (4)$$

where $\rho = p\epsilon$ is the product between the average probability p that an offspring will survive and reproduce and the average amount of energy ϵ that an individual on an unexploited resource allocates to reproduction. With ϵ amounts of energy allocated to reproduction ϵ/q offspring can be produced. Thus, the organism can choose to allocate resource either to the competitive quality q or to the population dynamic growth rate λ . In this paper I assume that the selection pressure on the three traits w , B , and T_p is similar so that I can model the evolutionary changes in the three traits by considering only the evolutionary changes in q . I will also assume that there are no evolutionary changes in p and ϵ so that the evolutionary changes in λ_m can be described by the evolutionary changes in q .

2.1 Evolutionary changes

To describe the evolutionary changes mathematically note, from eqns 4 and 3, that the population dynamic growth rate of the t th generation is

$$\lambda_t = \rho N_t^{-\gamma} / q_t \quad (5)$$

assuming that ρ is constant. The action of density dependent competitive interactions implies that the density regulation parameter γ is different for the different individuals in the population. The differentiation in γ arises because the competitively superior individuals dominate the competitively inferior individuals during competitive encounters. The competitively superior individuals will thus experience a lower cost of interference than the competitively inferior individuals. If we assume that the individuals encounter one another at random, the cost per unit interference (μ) of a given individual can be defined as a function of the competitive quality of that individual relative to the average competitive quality in the population. Hence, the density regulation parameter of the i th variant is $\gamma_i = \gamma_\alpha + \gamma_\iota(\mu + \Delta\mu_i)$, where $\Delta\mu_i = \mu_i - \mu$. If we assume that the cost of interference is a linear function on logarithmic scale it follows that $\Delta\mu_i = \psi(\ln q - \ln q_i)$, where ψ is the within population slope between the cost of interference competition and \ln competitive quality. Thus, the growth rate of the i th variant is

$$\begin{aligned} \lambda_{i,t} &= \rho N_t^{-\gamma_i} / q_{t,i} \quad (6) \\ &= \lambda_t (\rho N_t^{-\gamma_i} / q_{t,i}) / (\rho N_t^{-\gamma} / q_t) \\ &= \lambda_t (q_t / q_{i,t}) N_t^{\gamma - \gamma_i} \\ &= \lambda_t (q_t / q_{i,t}) N_t^{\gamma_\iota \psi (\ln q_{i,t} - \ln q_t)} \end{aligned}$$

which can be rearranged to

$$\lambda_{i,t} = \lambda_t e^{(\gamma_\iota \psi \ln N_t - 1)(\ln q_{i,t} - \ln q_t)} \quad (7)$$

Differentiating with respect to $\ln q_{i,t}$, and letting $q_{i,t} \rightarrow q_t$ the per generation selection gradient on $\ln q$ is

$$\partial \lambda_{i,t} / \partial \ln q_{i,t} |_{q_{i,t}=q_t} = \lambda_t (\gamma_\iota \psi \ln N_t - 1) \quad (8)$$

where $|_{q_{i,t}=q_t}$ indicates that the derivative is to be taken at the limit $q_{i,t} = q_t$. Solving for the evolutionary equilibrium we find that $N^{**} = e^{1/\gamma_\iota \psi}$, with $**$ denoting the evolutionary equilibrium. Combining this abundance with eqns 4 and 3 it follows that the evolutionary equilibrium is characterised as

$$\begin{aligned} \lambda &= \lambda_m^{**} N^{**-\gamma} = 1 \quad (9) \\ N^{**} &= \sqrt[\gamma]{\lambda_m^{**}} \end{aligned}$$

$$\begin{aligned} N^{**} &= e^{1/\gamma_\iota \psi} \\ \lambda_m^{**} &= e^{\gamma/\gamma_\iota \psi} \\ q^{**} &= \rho e^{-\gamma/\gamma_\iota \psi} \end{aligned}$$

To predict the per generation change in the growth rate λ_m , let $\sigma^2 = \hat{\sigma}^2 + \tilde{\sigma}^2$ represents the potential by which $\ln q$ responds to selection, where $\hat{\sigma}^2$ is the additive genetic variance in the Malthusian parameter $r = \ln \lambda_m$ and $\ln q$ (this variance is the same in the model because of the energetic trade-off between r and q) and $\tilde{\sigma}^2$ is a plastic response of inherited environmental effects. Assume also that the plastic response to selection can be modelled as the response of quantitative genetics, which implies that it operates by adjusting the phenotype of the offspring to the selection pressure experienced by the parents. Following Robertson (1968), Charlesworth (1990), Iwasa et al. (1991), Taper and Case (1992), Abrams et al. (1993), and Taylor (1996) the per generation change in $\ln q$ can then be approximated as

$$\begin{aligned} \Delta \ln q_t &= \frac{\sigma^2}{\lambda_t} \frac{\partial \lambda_{i,t}}{\partial \ln q_{i,t}} \Big|_{q_{i,t}=q_t} \quad (10) \\ &= \sigma^2 (\gamma_\iota \psi \ln N_t - 1) \end{aligned}$$

From eqn 4 note that $r = \ln \lambda_m = \ln \rho - \ln q$. Hence, with a constant ρ , σ^2 denotes also the potential response of the growth rate r to natural selection.

From eqn 10 and $\ln q_t = \ln q_{t-1} + \Delta \ln q_{t-1}$ the average quality at time t is

$$q_t = q_{t-1} N_{t-1}^{\gamma_\iota \psi \sigma^2} e^{-\sigma^2} \quad (11)$$

and from eqns 4 and 11, the maximal growth rate in generation t is

$$\begin{aligned} \lambda_{m,t} &= \rho N_{t-1}^{-\gamma_\iota \psi \sigma^2} e^{\sigma^2} / q_{t-1} \quad (12) \\ &= \lambda_{m,t-1} N_{t-1}^{-\gamma_q} e^{\sigma^2} \end{aligned}$$

with $\gamma_q = \gamma_\iota \psi \sigma^2$. Then, from eqns 3 and 12, the population dynamic equation with density dependent selection on competitive quality is

$$\begin{aligned} \lambda_{m,t} &= \lambda_{m,t-1} N_{t-1}^{-\gamma_q} e^{\sigma^2} \quad (13) \\ N_{t+1} &= N_t \lambda_{m,t} N_t^{-\gamma} \end{aligned}$$

This model reduces into the classical model of eqn 3 when $\sigma^2 = 0$, and it treats the growth rate as an initial condition, which contrasts to classical models where the growth rate is a fixed parameter.

Density independent environments

For eqn 13 it is the selection pressure of density dependent competitive interactions that generates the evolutionary balance between the competitive traits, the population dynamic growth rate, and the carrying capacity. This balance is lost if density regulation is absent. In this case $\gamma = 0$ and $\gamma_q = 0$ and, thus, from eqns 13 and 11 we obtain

$$\begin{aligned} q_t &= q_0 e^{-\sigma^2 t} \\ \lambda_{m,t} &= \lambda_{m,0} e^{\sigma^2 t} \\ N_t &= N_0 \lambda_{m,0}^t e^{\sigma^2 \sum_{\tau=0}^t \tau} \end{aligned} \quad (14)$$

a situation with a geometrical decline in competitive quality, a geometrical increase in the maximal growth rate, and a hyper-geometrical increase in the abundance.

To deduce the law behind eqn 14 recall that $r = \ln \lambda_m$, where r is the Malthusian parameter. Hence, from eqn 14 we obtain

$$\begin{aligned} r_t &= r_0 + \sigma^2 t \\ dr/dt &= \sigma^2 \end{aligned} \quad (15)$$

For the case of a constant environment and $\tilde{\sigma}^2 = 0$, this equation is Fisher's fundamental theorem of natural selection (Fisher 1930) that predicts that the partial increase in the Malthusian parameter due to natural selection is equal to the additive genetic variance in that parameter.

The observation that the fundamental theorem of natural selection may form the basis of population dynamics was probably first noted by Ginzburg (1980). From eqns 14 and 15 we note that the fundamental theorem and the law of hyper-geometrical increase reduces into the Malthusian law of geometrical increase (Malthus 1798) in the absence of both additive genetic variation and inherited environmental effects. A more detailed description of the relationship between the fundamental theorem and the selection pressure of density dependent competitive interactions is given by Witting (2000).

3 Model behaviour

The model with no selection [eqn 3] returns monotonically to the equilibrium when $0 < \gamma \leq 1$, and it oscillates with a period of two generations when $\gamma > 1$. The oscillation is damped for $1 < \gamma < 2$, stable for $\gamma = 2$, and repelling or exploding for $\gamma > 2$. This contrasts to the model with density dependent selection

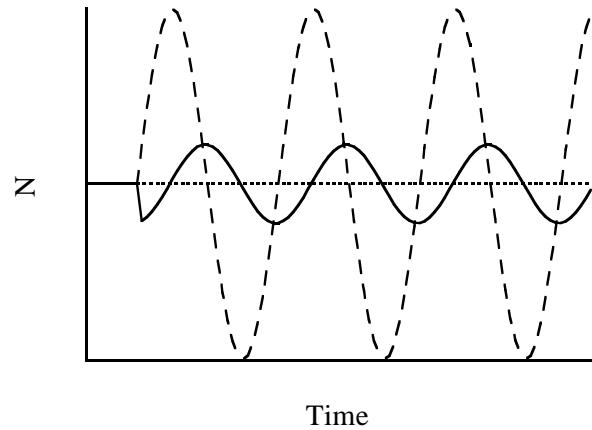


Figure 1: Projection of eqn 13. The solid curve is the population abundance, the dashed curve the population equilibrium, and the dotted line the evolutionary equilibrium.

[eqn 13], where the dynamics generally is cyclic with periods of more than two generations. The ultimate factor behind the cyclic dynamics of eqn 13 is the selection pressure of density dependent competitive interactions, but the proximate factor is probably best understood in terms of a cyclic change in population equilibrium. For the model with selection the population equilibrium $N^* = \sqrt[3]{\lambda_m}$ is defined as the equilibrium for a given λ_m and, thus, with cyclic and selection induced changes in λ_m there is a cycle in the population equilibrium. The evolutionary population equilibrium is the population equilibrium defined by the maximal growth rate at the evolutionary equilibrium, i.e., $N^{**} = \sqrt[3]{\lambda_m^{**}}$, with double star denoting evolutionary equilibrium.

The connection between the dynamics of the population equilibrium and the dynamics of the abundance is illustrated in Fig. 1. Following a downward perturbation from the evolutionary equilibrium the density regulation is relaxed so that the reproductive rate rises and the abundance increases toward the evolutionary equilibrium. Associated with the relaxed density regulation there is a decline in the level of interference competition and this decline selects for an increase in the population dynamic growth rate at the cost of the competitive traits. Thus, when the abundance has reached the evolutionary equilibrium the population will no longer be in population equilibrium at that equilibrium because the increased growth rate has induced an increase in the population equilibrium. Consequently, the population will increase beyond the evolutionary equilibrium heading toward the new population equilibrium. As illustrated in Fig. 1, the population will never reach this equilibrium because as soon as the abundance increases

beyond the evolutionary equilibrium the competitive interactions will select for an increase in the competitive traits at the cost of the population dynamic growth rate. Thus, the population equilibrium evolves downward toward the increasing population.

At some intermediate abundance the downward evolving equilibrium and the upward increasing population will intercept, and the population will be in population equilibrium with no change in abundance. At this equilibrium the abundance is at its peak, and selection for competitive quality is at its strongest. Hence, the population equilibrium will continue to evolve downward with the result that the abundance will decline in order to keep up with the downward evolving equilibrium. In Fig. 1 the abundance cannot keep up with the equilibrium, and consequently the distance between the abundance and the equilibrium increases until the population intercepts the evolutionary equilibrium with the fastest decline in abundance that the population encounters during the cycle. Having crossed the evolutionary equilibrium selection reverses, so that energy is allocated from the competitive traits to the population dynamic growth rate, and the population equilibrium evolves upward toward the declining population. At some intermediate abundance the declining population intercepts the upward evolving equilibrium, the rate of change in abundance changes from negative to positive, and the next period begins.

The dynamics of eqn 13 is analysed in Appendix A. The dynamics is generally cyclic with an either stable or unstable equilibrium. The equilibrium is unstable when $\gamma_q \geq \gamma$, or when $\gamma \geq 2$ and $\gamma_q \leq \gamma - 4$. The dynamics associated with the latter of these two criteria is given primarily by the over-compensatory density regulation that induces the oscillatory dynamics of eqn 3. The result is oscillatory dynamics with a period of two generations. The dynamics associated with the former criterion is caused primarily by the evolutionary modulation of life history characters. This dynamics is cyclic with a highly variable period. The cycles are damped when $\gamma_q < \gamma$, stable when $\gamma_q = \gamma$, and repelling or exploding when $\gamma_q > \gamma$.

When the cyclic dynamics generated by density dependent selection is stable, i.e., when $\gamma_q = \gamma$, the period is determined by the γ parameter. In this case the period declines from an infinite number of generations for $\gamma = 0$ to two generations for $\gamma = 4$. From eqn 12 we have that $\gamma_q = \gamma_i \psi \sigma^2$. Thus, the population goes through faster cycles *i*) when the density dependence (γ & γ_q) increases, *ii*) when the slope between the cost of interference and \ln competitive quality (ψ) increases, and *iii*) when the additive genetic variance ($\hat{\sigma}^2$) and/or

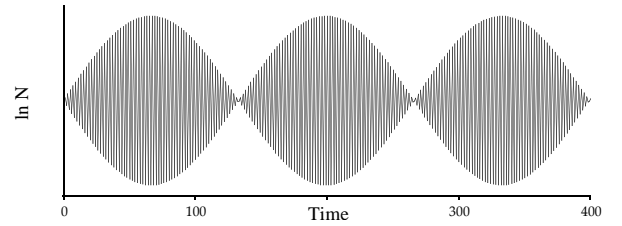


Figure 2: A projection of eqn 13 showing the occurrence of a cycle in the amplitude of the population cycle. In the figure, the population cycle has a period of ≈ 2 generations, and the amplitude has a period of ≈ 133 generations.

the plastic response ($\hat{\sigma}^2$) to selection increase.

In the case where $\gamma_q = \gamma$ and $3.5 \leq \gamma \leq 4$ there is a cycle in the amplitude of the population cycle, as it is illustrated in Fig. 2. On the continuum from $\gamma = 3.5$ to $\gamma = 4$ the population cycle is dominated increasingly by the cyclic change in amplitude. As $\gamma \rightarrow 4$ the period of the amplitude cycle continues to increase, from ≈ 10 generations for $\gamma = 3.9$ to an infinite number of generations at the limit $\gamma = 4$. Beyond $\gamma = 4$ the population period is two generations, while the amplitude increases geometrically.

Associated with the cyclic dynamics of eqn 13 there are cyclic changes in both the maximal population dynamic growth rate and the competitive traits. These changes are described by eqns 13 and 11, which predict that the maximal growth rate increases whenever the population abundance is below the abundance of the evolutionary equilibrium, and that the growth rate declines when the abundance is above the abundance at equilibrium. The reverse is true for competitive traits like the body mass, the metabolic rate, and the pre-reproductive period, which are treated collectively by the competitive quality q . During the population cycle the competitive traits will have their maximal values when the abundance is crossing the population equilibrium from above, and they will take their minimal values when the abundance is crossing the equilibrium from below.

4 Discussion

Even though the deduction presented here is a bit simplified, it clearly shows that density dependent competitive interactions selects for density dependent changes in individual quality and for cyclic population dynamics; a deduction that has been missing since the first selection hypothesis was proposed by Chitty (1960). In the presented model the population dynamic growth

rate is treated as an initial condition and not as a parameter, as it usually is the case in population dynamic equations. And the density independent fundament is Fisher's (1930) fundamental theorem of natural selection, which predicts a hyper-geometrical increase in the abundance instead of the geometrical increase of the Malthusian law (Malthus 1798). The model may thus be seen as a variant of the inertia growth initially proposed by Ginzburg (1980, 1986) [see also related discussion in Clark (1971) and Innis (1972)]. Where Ginzburg (1998) more recently has developed inertia growth into the hypothesis of maternal effects, my study suggests that it may instead be understood by a general law of selection by density dependent competitive interactions.

It is often argued that a successful theory of cyclic population dynamics needs to explain a set of observations (e.g., Krebs and Myers 1974; Stenseth 1985). These observations include (i) that cyclic dynamics is more common and that it have larger amplitudes toward arctic regions, (ii) that the period of cyclic dynamics is more fixed than the amplitude, (iii) the actual periods of the cyclic dynamics, and (iv) the co-occurrence of cyclic population dynamics and cyclic life histories. As described in the following three subsections these observations can be explained by the mechanism of selection by density dependent competitive interactions. The resemblance between the predicted dynamics and the dynamics of animal species does not imply that the proposed mechanism necessarily applies to natural populations. In line with other recent studies (e.g., Ginzburg and Taneyhill 1994; Inchausti and Ginzburg 1998), the resemblance shows only that the hypothesis is a plausible mechanism. To determine the actual causes in natural populations we need empirical evidence, but empirical studies have so far failed to determine the true cause for cyclic population dynamics (e.g., Krebs 1996).

4.1 Neutral stability

Like the dynamics of many other models, the cycles of eqn 13 are neutrally stable in the sense that their amplitude is given by initial conditions even though the period is fixed and given by the particular values of the density regulation parameters. This implies that the amplitude of the cycle will increase on a cline from a stable to a fluctuating environment. One such cline is from temperate areas toward the polar regions where an increased frequency and severity of unfavourable climatic conditions occur. On this cline it is generally observed that population cycles tend to vanish toward

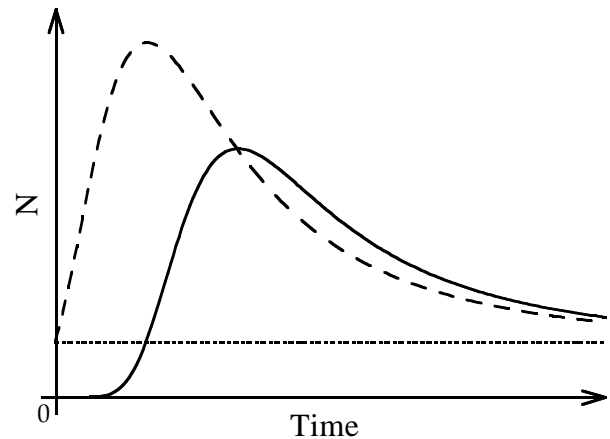


Figure 3: A projection of eqn 13 illustrating the dynamics of an introduced species with strongly damped dynamics ($\gamma = 0.07$ & $\gamma_q = 0.01\gamma$). At time zero the species is introduced at an abundance that is 10^{-7} times smaller than the abundance at the evolutionary equilibrium. The solid curve is the population abundance, the dashed curve the population equilibrium, and the dotted line the evolutionary equilibrium.

the south while they are more pronounced with progressively larger amplitudes toward Arctic regions (e.g., Howell 1923; Hansson 1971; Hansson and Henttonen 1985; Stenseth et al. 1985; Hansson 1987; Hanski et al. 1991; Akçakaya 1992; Bjørnstad et al. 1995).

As the amplitude of neutrally stable dynamics depends on the magnitude of the environmental perturbation a special type of dynamics can arise from extreme perturbations of equilibria with strongly damped dynamics. In Fig. 3 I have used eqn 13 to simulate the dynamics following a perturbation to an extraordinarily low abundance. At this abundance interference competition is almost absent and that there is strong selection for an increase in the population dynamic growth rate. The result is a lack phase with only a marginal increase before the evolutionary increase in population dynamic growth rate causes the population to explode to an extreme abundance. The peak abundance is associated with high levels of interference competition where energy is selected from demographic traits into competitive traits causing the population to decline slowly to a normal and stable abundance at the evolutionary equilibrium. This type of dynamics with a single, or a few, over-shoots is often observed when humans release exotic species into new areas (e.g., Elton 1927; Adam et al. 1993).

Species	n	\bar{N}	P	c	$\gamma \pm \text{SE}$	$\sigma^2 \pm \text{SE}$	N^{**}
Douglas-fir tussock moth	10	4.1	9.0	0.96	0.52 ± 0.17	0.77 ± 0.41	4.39
Fall webworm	22	10.2	7.0	0.82	0.69 ± 0.12	1.54 ± 0.44	9.83
Larch budmoth	38	2.2	9.3	0.96	0.53 ± 0.07	0.51 ± 0.28	2.62
Larch cone fly	10	945	4.5	0.32	1.68 ± 0.09	11.5 ± 0.88	950
Nun moth	42	2600	4.1	0.55	1.28 ± 0.05	10.0 ± 0.61	2630
Pine looper moth	50	0.66	4.9	0.62	1.30 ± 0.14	-0.48 ± 0.19	0.69
Southern pine beetle	30	700	7.7	0.94	0.42 ± 0.04	2.67 ± 0.39	605
Spruce budworm	28	1.5	21	0.94	0.28 ± 0.13	0.16 ± 0.31	1.76
Wasp spp.	25	10.2	2.2	-0.57	2.89 ± 0.31	6.59 ± 1.05	9.78

Table 1: The parameter estimates for eqn 13 for nine insect species (assuming $\gamma_q = \gamma$). n is the number of years with abundance estimates, \bar{N} the geometric mean in the data, P the average period in the data, c the correlation coefficient of the regression [eqn 27] that was used to estimate the two parameters γ and σ^2 , and N^{**} the estimated population equilibrium. The estimation procedure is described in Appendix B. Data from Royama (1984), Turchin et al. (1991), Dahlsten et al. (1990), Baltensweiler and Fischlin (1988), Morris (1964), Bejer (1988), Barbour (1990), Roques (1988), and Southwood (1967).

4.2 The period

Forest insects are known for their pronounced population cycles. In the literature I found nine species, mainly Lepidoptera, with yearly non-overlapping generations, cyclic dynamics, and yearly abundance estimates for periods longer than ten years. Assuming that the cycles are stable, i.e., that $\gamma_q = \gamma$, the parameters of eqn 13 can be estimated from such time series by regression eqn 27 in Appendix B. Table 1 lists the estimated models. The estimated evolutionary equilibria (N^{**}) resemble the geometric mean of the abundance in the time series (N_G), a result that supports the assumption of linearity on logarithmic scale.

As mentioned in Section 2, the period in the dynamics of eqn 13 is given by the γ parameter, and in Fig. 4 this relationship is shown by the curve while the diamonds represent the nine species. Probably the best documented cycle is found in the larch budmoth that “goes through 10000-fold changes in density during its very regular [8.24 \pm 0.27 (SE) years] cycle in the Alps” (Berryman 1996). The data on one population of this species are shown by the diamonds in Fig. 5, where the curve represents a projection of the model in eqn 13. The resemblance between theory and data in Figs. 4 and 5 is reasonable suggesting that selection by density dependent competitive interactions may play a role in the cyclic dynamics of forest insects. But other hypotheses, like the maternal effect hypothesis (Ginzburg and Taneyhill 1994, 1995; Ginzburg 1998), have produced comparable fits and the resemblance may be a more general result of second order population dynamic equations.

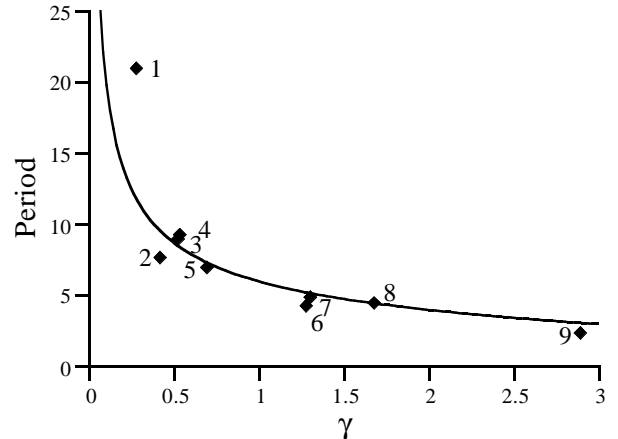


Figure 4: The period of the population cycle in generations against the γ parameter (assuming $\gamma_q = \gamma$). The curve is defined by eqn 13, and the numbered diamonds represent the following species: 1: Spruce budworm. 2: Southern pine beetle. 3: Douglas-fir tussock moth. 4: Larch budmoth. 5: Fall webworm. 6: Nun moth. 7: Pine looper moth. 8: Larch cone fly. 9: Wasp spp.

4.3 Life history cycles

Selection by density dependent competitive interactions suggests that population cycles are associated with cyclic changes in life history parameters like body mass, metabolic rate, and age of reproduction. Such changes have often been found in small rodents with cyclic dynamics (Krebs and Myers 1974; Krebs 1978; Boonstra and Krebs 1979; Stenseth 1982; Stenseth and Ims 1993; Lidicker and Ostfeld 1991). It is observed that rodents are small, non-aggressive, and that they have a high rate of reproduction when the abundance is low and in-

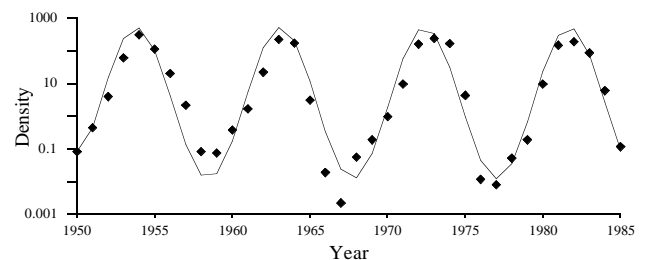


Figure 5: The curve is a projection of eqn 13 and the diamonds the yearly abundance of the larch budmoth in the Upper Engadine valley from 1950 to 1985. The γ and σ^2 parameters of eqn 13 are adjusted from those given in Table 1 to $\gamma = 0.44$ and $\sigma^2 = 0.42$ so that the equilibrium is the same and the period resembles the observed period. Data from Baltensweiler and Fischlin (1988).

creasing. When instead the abundance is high and declining the rodents are aggressive, 20-30% larger than at low abundance, and they have delayed reproduction and a low reproductive rate.

The cyclic changes in life history characters are also found in other species. Murdoch and McCauley (1985) observed cyclic dynamics in a *Daphnia* population, and Fig. 6 shows the average body mass and the population abundance against time. It is apparent that the body mass cycles, and that this cycle is lagging behind the cycle in abundance, as predicted by the dynamics of eqn 13. From eqn 11 the predicted relation between body mass (w) and population abundance is

$$\ln w_t - \ln w_{t-1} = \gamma_q \ln N_{t-1} + c \quad (16)$$

where c is a constant. This relation is significantly present in the *Daphnia* population (correlation coefficient = 0.49, $n = 26$, $\gamma_q = 1.7 \pm 0.5(\text{SE})$, and $c = -0.55 \pm 0.36$).

Some studies have reported no correlation between the body mass and the phase of the abundance cycle (Ferns 1979; Myllymäki 1977), but the widespread observation of phase related life history cycles has been seen as essential for a mechanistic understanding of cyclic population dynamics (e.g., Chitty 1987, 1996; Krebs 1996). As the body mass tends to be largest at the peak abundance the life history cycles cannot be explained as simple reflections of density dependent changes in the environment. The hypothesis of density dependent competitive interactions incorporates the cyclic changes in life history parameters as part of the causal agent that is responsible for the delayed density dependence that is so essential for the generation of cyclic dynamics.

Selection by density dependent competitive interactions suggests that selection is strongest when the abundance is at peak and bottom densities, while selection can be absent at intermediate densities. For a rodent cycle Krebs et al. (1973) found that selection was strongest at the peak abundance where the frequency of the LAPs allele changed from ≈ 0.70 to ≈ 0.40 in less than one generation. And for the oak leafroller moth Simchuk et al. (1999) found that trends at the Est-4 and Pts-4 loci were directly related to the population dynamics. They also found density dependent changes in the selection pressure on body mass, with larger females being selected prior to the population decline, and smaller females being selected during the outbreak phase of the cycle.

Although density dependent selection may be responsible for the cyclic changes in body mass in some cases,

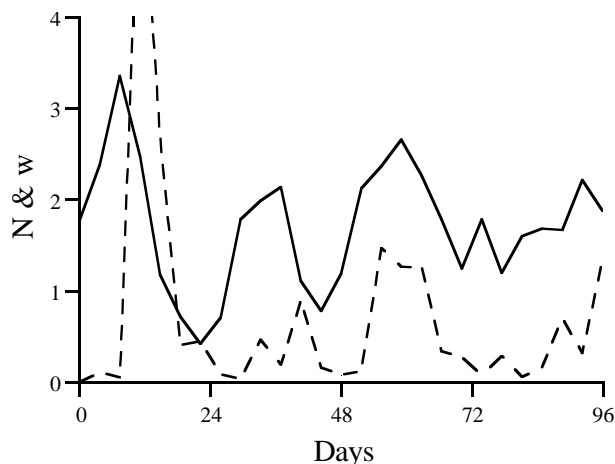


Figure 6: The dynamics in the abundance (N) and body mass (w) of a *Daphnia* population against time. The solid curve is the population abundance (10^5 *Daphnia* per m^{-3}) and the dashed curve the body mass (μg dry weight). Data from Fig. 3 in Murdoch and McCauley (1985).

in other situations cyclic body masses may be a consequence of cyclic dynamics and not necessarily part of the cause that drives the cycle (Lidicker and Ostfeld 1991). This can be the case if changes in foraging time and reproductive effort cause phase related changes in body mass (Oksanen and Lundberg 1995). Or if phase related changes in body mass are induced by dynamic energy allocation in fluctuating environments, with fluctuating environments including population dynamic cycles (Oli 1999).

4.4 Allometries of population dynamics

The allometric relationship between body mass and respectively the population abundance and the period of the cyclic dynamics are other relations that a successful population dynamic theory needs to explain. For species outside competitive guilds, the population abundance is approximately proportional to the negative 3/4 power of body mass (Damuth 1981, 1987; Nee et al. 1991), and the Calder hypothesis suggests that the period of population cycles scale to the 1/4 power of body mass (Calder 1983, 1984; Peters 1983; Peterson et al. 1984; Krukoniš and Schaffer 1991). The population abundance allometry has been shown to apply to a variety of taxa, and the Calder hypothesis is confirmed for terrestrial homiotherms.

The exponents of the abundance and population cycle allometries follow from first principles of the proposed density dependent ecology. The model of eqn 13 is part of a larger theory where the across species ex-

ponents of the lifespan and population abundance allometries are deduced from the density dependent constraints that competitive interactions and foraging self inhibition places on the foraging process in mobile organisms (Witting 1995, 1998). Lifespan (T) is predicted to scale as $T \propto w^{1/2d}$, and the abundance as $N^* \propto w^{(1-2d)/2d}$, where w is body mass and d the number of dimensions in which the organism forages. Thus, for terrestrial organisms, which are likely to forage in two dimensions, lifespan is expected to scale to the 1/4 power of body mass and the population abundance to the negative 3/4 power of body mass. As the period of eqn 13 is given in generations it follows that the period in astronomical time is proportional to the 1/4 power of body mass, as suggested empirically by the Calder hypothesis.

For organisms that forage in three dimensions the predicted exponents are 1/6 and 5/6 instead of 1/4 and 3/4. The dependence of the allometric exponents upon the dimensionality of the foraging behaviour is empirically confirmed for the lifespan exponent, which is 0.25 ± 0.04 (SE) among 195 species of terrestrial mammals while it is 0.16 ± 0.02 (SE) among 40 species of pelagic mammals expected to forage in three spatial dimension (Witting 1995, 1998). But whether the exponents of the abundance and population cycle allometries for pelagic species are 5/6 and 1/6 instead of 3/4 and 1/4 remain to be shown.

References

- Abrams P. A., Harada Y., Matsuda H. (1993). On the relationship between quantitative genetic and ESS models. *Evolution* 47:982–985.
- Adam K. D., King C. M., Köhler W. H. (1993). Potential ecological effects of escaped transgenic animals: lessons from past biological invasions. In: Wöhrmann K. Tomiuk J. (eds). *Transgenic Organisms*: Birkhäuser Verlag, Basel, pp 153–173.
- Akçakaya H. R. (1992). Population cycles of mammals: evidence for a ratio-dependent predation hypothesis. *Ecol. Monogr.* 62:119–142.
- Axelrod R. Hamilton W. D. (1981). The evolution of cooperation. *Science* 211:1390–1396.
- Baltensweiler W. Fischlin A. (1988). The larch budmoth in the alps. In: Berryman A. A. (ed). *Dynamics of forest insect populations. Patterns, causes, implications*: Plenum Press, New York, pp 331–351.
- Barbour D. A. (1990). Synchronous fluctuations in spatially separated populations of cyclic forest insects. In: Watt A. D., Leather S. R., Hunter M. D., Kidd N. A. C. (eds). *Population dynamics of forest insects*: Intercept, Andover, Hampshire, pp 339–346.
- Batzli G. O. (1996). Population cycles revisited. *Trends Ecol. Evol.* 11:488–489.
- Bejer B. (1988). The nun moth in European spruce forests. In: Berryman A. A. (ed). *Dynamics of forest insect populations. Patterns, causes, implications*: Plenum Press, New York, pp 211–231.
- Berryman A. A. (1996). What causes population cycles of forest lepidoptera? *Trends Ecol. Evol.* 11:28–32.
- Bjørnstad O. N., Falck W., Stenseth N. C. (1995). A geographic gradient in small rodent density fluctuations: a statistical modelling approach. *Proc. R. Soc. Lond. B.* 262:127–133.
- Boonstra R. (1994). Population cycles in microtines: the senescence hypothesis. *Evol. Ecol.* 8:196–216.
- Boonstra R. Boag P. T. (1987). A test of the Chitty hypothesis: inheritance of life-history traits in meadow voles *Microtus pennsylvanicus*. *Evolution* 41:929–947.
- Boonstra R. Hochachka W. M. (1997). Maternal effects and additive genetic inheritance in the collared lemming *Dicrostonyx groenlandicus*. *Evol. Ecol.* 11:169–182.
- Boonstra R. Krebs C. J. (1979). Viability of large- and small-sized adults in fluctuating vole populations. *Ecology* 60:567–573.
- Bulmer M. (1994). *Theoretical evolutionary ecology*. Sinauer Associates Publishers, Massachusetts.
- Calder W. A. I. (1983). An allometric approach to population cycles of mammals. *J. theor. Biol.* 100:275–282.
- Calder W. A. I. (1984). *Size, function, and life history*. Harvard University Press, Cambridge.
- Charlesworth B. (1990). Optimization models, quantitative genetics, and mutation. *Evolution* 44:520–538.
- Charlesworth B. (1994). *Evolution in age-structured populations*. 2nd edn. Cambridge University Press, Cambridge.
- Charnov E. L. (1993). *Life history invariants. Some explorations of symmetry in evolutionary ecology*. Oxford University Press, New York.
- Charnov E. L. Finnerty J. (1980). Vole population cycles: a case for kin-selection? *Oecologia* 45:1–2.
- Chitty D. (1960). Population processes in the voles and their relevance to general theory. *Can. J. Zool.* 38:99–113.
- Chitty D. (1967). The natural selection of self-regulatory behaviour in animal populations. *Proc. Ecol. Soc. Aust.* 2:51–78.
- Chitty D. (1987). Social and local environments of the vole *Microtus townsendii*. *Can. J. Zool.* 65:2555–2566.
- Chitty D. (1996). *Do lemmings commit suicide? Beautiful hypotheses and ugly facts*. Oxford University Press, New York.
- Christian J. (1950). The adreno-pituitary system and population cycles in mammals. *J. Mamm.* 31:247–259.
- Christiansen F. B. (1991). On conditions for evolutionary stability for a continuously varying character. *Am. Nat.* 138:37–50.
- Clark J. (1971). The second derivative and population mod-

- eling. Ecology 52:606–613.
- Costantino R. F., Cushing J. M., Dennis B., Desharnais R. A. (1995). Experimentally induced transitions in the dynamic behavior of insect populations. Nature 375:227–230.
- Dahlsten D. L., Rowney D. L., Copper W. A., Tait S. M., Wenz J. M. (1990). Long-term population studies of the douglas-fir tussock moth in california. In: Watt A. D., Leather S. R., Hunter M. D., Kidd N. A. C. (eds). Population dynamics of forest insects: Intercept, Andover, Hampshire, pp 45–58.
- Damuth J. (1981). Population density and body size in mammals. Nature 290:699–700.
- Damuth J. (1987). Interspecific allometry of population density in mammals and other animals: the independence of body mass and population energy-use. Biol. J. Linn. Soc. 31:193–246.
- Dekker H. (1975). A simple mathematical model of rodent population cycles. J. Math. Biol. 2:57–67.
- Desharnais R. A., Liu L. (1987). Stable demographic limit cycles in laboratory populations of *tribolium castaneum*. J. Anim. Ecol. 56:885–906.
- Dieckmann U. (1997). Can adaptive dynamics invade? Trends Ecol. Evol. 12:128–131.
- Dugatkin L. A., Reeve H. K. (1998). Game theory and animal behavior. Oxford University Press, Oxford.
- Elton C. (1927). Animal Ecology. Sidgwick & Jackson, London.
- Eshel I. (1983). Evolutionary and continuous stability. J. theor. Biol. 103:99–111.
- Eshel I., Motro U., Sansone E. (1997). Continuous stability and evolutionary convergence. J. theor. Biol. 185:333–343.
- Ferns P. N. (1979). Growth, reproduction and residency in a declining population of *Microtus agrestis*. J. Anim. Ecol. 48:739–758.
- Fisher R. A. (1930). The genetical theory of natural selection. Clarendon, Oxford.
- Geritz S. A. H., Kisdi É., MeszÉna G., Metz J. A. J. (1998). Evolutionary singular strategies and the adaptive growth and branching of the evolutionary tree. Evol. Ecol. 12:35–57.
- Geritz S. A. H., Metz J. A. J., Kisdi É., MeszÉna G. (1997). Dynamics of adaptation and evolutionary branching. Phys. Rev. Lett. 78:2024–2027.
- Ginzburg L. R. (1980). Ecological implications of natural selection. In: Barigozzi C. (ed). Vito Volterra symposium on mathematical models in biology. Lecture notes in biomathematics, Vol. 39: Springer-Verlag, Berlin, pp 171–183.
- Ginzburg L. R. (1986). The theory of population dynamics: I. Back to first principles. J. theor. Biol. 122:385–399.
- Ginzburg L. R. (1998). Inertial growth. Population dynamics based on maternal effects. In: Mousseau T. A., Fox C. W. (eds). Maternal effects as adaptations: Oxford University Press, New York, pp 42–53.
- Ginzburg L. R., Taneyhill D. E. (1994). Population cycles of forest lepidoptera: a maternal effect hypothesis. J. Anim. Ecol. 63:79–92.
- Ginzburg L. R., Taneyhill D. E. (1995). Higher growth rate implies shorter cycle, whatever the cause: a reply to berryman. J. Anim. Ecol. 64:294–295.
- Grenfell B. T., Price O. F., Albon S. D., Clutton-Brock T. H. (1992). Overcompensation and population cycles in an ungulate. Nature 355:823–826.
- Gurney W. S. C., Blythe S. P., Nisbet R. M. (1980). Nicholson’s blowflies revisited. Nature 287:17–21.
- Hanski I., Hansson L., Henttonen H. (1991). Specialist predators, generalist predators, and the microtine rodent cycle. J. Anim. Ecol. 60:353–367.
- Hanski I., Korpimäki E. (1995). Microtine rodent dynamics in northern europe: Parameterized models for the predator-prey interaction. Ecology 76:840–850.
- Hanski I., Turchin P., Korpimäki E., Henttonen H. (1993). Population oscillations of boreal rodents: regulation by mustelid predators leads to chaos. Nature 364:232–235.
- Hansson L. (1971). Small rodent food, feeding and population dynamics: a comparison between granivorous and herbivorous species in Scandinavia. Oikos 22:183–198.
- Hansson L. (1987). An interpretation of rodent dynamics as due to trophical interactions. Oikos 50:308–318.
- Hansson L., Henttonen H. (1985). Gradients in density variations of small rodents: the importance of latitude and snow cover. Oecologia 67:394–402.
- Hårdling R. (1999). Arms races, conflict costs and evolutionary dynamics. J. theor. Biol. 196:163–167.
- Hassell M. P., Lawton J. H., May R. M. (1976). Patterns of dynamical behavior in single species populations. J. Anim. Ecol. 45:471–486.
- Hofbauer, J. & Sigmund, K., eds (1998). Evolutionary games and population dynamics. Cambridge University Press, Cambridge.
- Howell A. B. (1923). Periodic fluctuations in the numbers of small mammals. J. Mamm. 4:149–155.
- Hunt F. (1982). Regulation of population cycles by genetic feedback. existence of period solutions of a mathematical model. J. Math. Biol. 13:271–282.
- Inchausti P., Ginzburg L. R. (1998). Small mammals cycles in northern Europe: patterns and evidence for a maternal effect hypothesis. J. Anim. Ecol. 67:180–194.
- Innis G. (1972). The second derivative and population modeling: another view. Ecology 53:720–723.
- Iwasa Y., Pomiankowski A., Nee S. (1991). The evolution of costly mate preferences. II. The “handicap” principle. Evolution 45:1431–1442.
- Jablonka E., Lamb M. J. (1989). The inheritance of acquired epigenetic variations. J. theor. Biol. 139:69–83.
- Jablonka E., Lamb M. J. (1998). Epigenetic inheritance in evolution. J. Evol. Biol. 11:159–183.
- Jedrzejewski W., Jedrzejewski B. (1996). Rodent cycles in

- relation to biomass and productivity of ground vegetation and predation in the Palearctic. *Acta Theriol.* 41:1–34.
- Kisdi É. (1999). Evolutionary branching under asymmetric competition. *J. theor. Biol.* 197:149–162.
- Kozłowski J. (1999). Adaptation: a life history perspective. *Oikos* 86:185–194.
- Krebs C. J. (1978). A review of the chitty hypothesis of population regulation. *Can. J. Zool.* 56:2464–2480.
- Krebs C. J. (1996). Population cycles revisited. *J. Mamm.* 77:8–24.
- Krebs C. J., Boutin S., Boonstra R., Sinclair A. R. E., Smith J. N. M., Dale M. R. T., Martin K., Turkington R. (1995). Impact of food and predation on the snowshoe hare cycle. *Science* 269:1112–1115.
- Krebs C. J., Gaines M. S., Keller B. L., Myers J. H., Tamarin R. H. (1973). Population cycles in small rodents. *Science* 179:35–41.
- Krebs C. J., Myers J. (1974). Population cycles in small mammals. *Ad. Ecol. Res.* 8:267–399.
- Krukoniš G., Schaffer W. M. (1991). Population cycles in mammals and birds: does periodicity scale with body size? *J. theor. Biol.* 148:469–493.
- Lachmann M., Jablonka E. (1996). The inheritance of phenotypes: an adaptation against fluctuating environments. *J. theor. Biol.* 181:1–9.
- Lidicker W. Z., Ostfeld R. S. (1991). Extra-large body size in California voles: Causes and fitness consequences. *Oikos* 61:108–121.
- Malthus T. R. (1798). *An essay on the principle of population.* Johnson, London.
- Matsuda H., Abrams P. A. (1994). Runaway evolution to self-extinction under asymmetrical competition. *Evolution* 48:1764–1772.
- May R. M., Oster G. F. (1976). Bifurcation and dynamic complexity in simple ecological models. *Amazoniana* 110:573–599.
- Maynard Smith J. (1982). *Evolution and the theory of games.* Cambridge University Press, Cambridge.
- Maynard Smith J., Brown R. L. W. (1986). Competition and body size. *Theor. Pop. Biol.* 30:166–179.
- Maynard Smith J., Price G. R. (1973). The logic of animal conflict. *Nature* 246:15–18.
- Metz J. A. J., Geritz S. A. H., Meszéna G., Jacobs F. J. A., vanHeerwaarden J. S. (1996). Adaptive dynamics, a geometrical study of the consequences of nearly faithful reproduction. In: vanStrien S. J., Verduyn Lunel S. M. (eds). *Stochastic and spatial structures of dynamical systems: North Holland, Amsterdam, The Netherlands*, pp 183–231.
- Metz J. A. J., Nisbet R. M., Geritz S. A. H. (1992). How should we define fitness? for general ecological scenarios? *Trends Ecol. Evol.* 7:198–202.
- Morris R. F. (1964). The value of historical data in population research, with particular reference to *hyphantria cunea druryi*. *Can. Entomol.* 96:356–368.
- Mueller L. D. (1997). Theoretical and empirical examination of density-dependent selection. *Ann. Rev. Ecol. Syst.* 28:269–288.
- Murdoch W. W., McCauley E. (1985). Three distinct types of dynamic behavior shown by a single planktonic system. *Nature* 316:628–630.
- Myllymäki A. (1977). Demographic mechanisms in the fluctuating populations of the field vole *macrotus agrestis*. *Oikos* 468–493:212–214.
- Nee S., Read A. F., Greenwood J. J. D., Harvey P. H. (1991). The relationship between abundance and body size in british birds. *Nature* 351:312–313.
- Norrdahl K. (1995). Population cycles in northern small mammals. *Biol. Rev.* 70:621–637.
- Oksanen L., Lundberg P. (1995). Optimization of reproductive effort and foraging time in mammals. The influence of resource level and predation risk. *Evol. Ecol.* 9:54–56.
- Oli M. K. (1999). The Chitty effect: A consequence of dynamic energy allocation in a fluctuating environment. *Theor. Pop. Biol.* 56:293–300.
- Peters R. H. (1983). *The ecological implication of body size.* Cambridge University Press, Cambridge.
- Peterson R. O., Page R. E., Dodge K. M. (1984). Wolves, moose, and the allometry of population cycles. *Science* 224:1350–1352.
- Price G. R. (1972). Fisher's "fundamental theorem" made clear. *Ann. Hum. Genet.* 36:129–140.
- Robertson A. (1968). The spectrum of genetic variation. In: Lewontin R. C. (ed). *Population Biology and Evolution: Syracuse University Press, New York*, pp 5–16.
- Roff D. A. (1992). *The evolution of life histories. Theory and analysis.* University of Chicago Press, New York.
- Roques A. (1988). The larch cone fly in the french alps. In: Berryman A. A. (ed). *Dynamics of forest insect populations. Patterns, causes, implications: Plenum Press, New York*, pp 1–28.
- Rossiter M. C. (1991). Environmentally-based maternal effects: a hidden force in insect population dynamics. *Oecologia* 87:288–294.
- Rossiter M. C. (1992). The impact of resource variation on population quality in herbivorous insects: a critical component of population dynamics. In: Hunter M. D., Ohgushi T., Price P. W. (eds). *Resource distribution and animal-plant interactions: Academic Press, New York*, pp 13–42.
- Rossiter M. C. (1994). Maternal effects hypothesis of herbivore outbreak. *BioSci.* 44:752–763.
- Rossiter M. C. (1996). Incidence and consequences of inherited environmental effects. *Ann. Rev. Ecol. Syst.* 27:451–476.
- Royama T. (1984). Population dynamics of the spruce budworm *choristoneura fumiferana*. *Ecol. Monogr.* 54:429–462.

- Sandefur J. T. (1990). Discrete dynamic systems: Theory and applications. Oxford University Press, Oxford.
- Simchuk A. P., Ivashov A. V., Companytsev V. A. (1999). Genetic patters as possible factors causing population cycles in oak leafroller moth, *Tortrix viridana* L. For. Ecol. Manage. 113:35–49.
- Southwood T. R. E. (1967). The interpretation of population change. J. Anim. Ecol. 36:519–529.
- Stearns S. C. (1992). The evolution of life histories. Oxford University Press, Oxford.
- Stenseth N. C. (1978). Demographic strategies in fluctuating populations of small rodents. Oecologia 33:149–172.
- Stenseth N. C. (1981). On chitty's theory for fluctuating population: the importance of genetic polymorphism in the generation of regular density cycles. J. theor. Biol. 90:9–36.
- Stenseth N. C. (1982). Causes and consequences of dispersal in small mammals. In: Swingland I. Greenwood P. (eds). The ecology of animal movement: Oxford University Press, Oxford, pp 62–101.
- Stenseth N. C. (1985). Mathematical models of microtine cycles: models and the real world. Acta Zool. Fenn. 173:7–12.
- Stenseth N. C. (1995). Snowshoe hare populations: Squeezed from below and above. Science 269:1061–1062.
- Stenseth N. C., Gustafsson T. O., Hansson L., Uglund K. I. (1985). On the evolution of reproductive rates in microtine rodents. Ecology 66:1795–1808.
- Stenseth, N. C. & Ims, R., eds (1993). The biology of lemmings. Academic Press, San Diego.
- Strogatz S. H. (1994). Nonlinear dynamics and chaos. Addison-Wesley, Reading, MA.
- Taper M. L. Case T. J. (1992). Models of character displacement and the theoretical robustness of taxon cycles. Evolution 46:317–333.
- Taylor P. D. (1989). Evolutionary stability in one-parameter models under weak selection. Theor. Pop. Biol. 36:125–143.
- Taylor P. D. (1996). The selection differential in quantitative genetics and ess models. Evolution 50:2106–2110.
- Thue Poulsen E. (1979). A model for population regulation with density- and frequency-dependent selection. J. Math. Biol. 8:325–343.
- Tuljapurkar S., Boe C., Wachter K. W. (1994). Nonlinear feedback dynamics in fisheries: Analysis of the Deriso-Schnute model. Can. J. Fish. Aquat. Sci. 51:1462–1473.
- Turchin P. (1990). Rarity of density dependence or population regulation with lags? Nature 344:660–663.
- Turchin P., Lorio J. P. L., Taylor A. D., Billings R. F. (1991). Why do populations of southern pine beetles (coleoptera: Scolytidae) fluctuate? Environ. Entomol. 20:401–409.
- Turchin P., Oksanen L., Ekerholm P., Oksanen T., Henttonen H. (2000). Are lemmings prey or predators? Nature 405:562–565.
- Turchin P. Taylor A. D. (1992). Complex dynamics in ecological time series. Ecology 73:289–305.
- Vega-Redondo F. (1996). Evolution, Games, and Economic Behaviour. Oxford University Press, Oxford.
- Vincent T. L. Brown J. S. (1988). The evolution of ess theory. Ann. Rev. Ecol. Syst. 19:423–443.
- Wellington W. G. (1965). Some maternal influences on progeny quality in the western tent caterpillar *malacosoma pluviale*. Can. Entomol. 97:1–14.
- Witteman G. J., Redfearn A., Pimm S. L. (1990). The extent of complex population changes in nature. Evol. Ecol. 4:173–183.
- Witting L. (1995). The body mass allometries as evolutionarily determined by the foraging of mobile organisms. J. theor. Biol. 177:129–137.
- Witting L. (1997). A general theory of evolution. By means of selection by density dependent competitive interactions. Peregrine Publisher, Århus, 330 pp, URL <http://mrLife.org>.
- Witting L. (1998). Body mass allometries caused by physiological or ecological constraints? Trends Ecol. Evol. 13:25.
- Witting L. (2000). Interference competition set limits to the fundamental theorem of natural selection. Acta Biotheor. 48:107–120.

A Stability and dynamic behaviour

The stability of the evolutionary equilibrium and the dynamics that follow from a perturbation of that equilibrium is analysed in this section. Let $r = \ln \lambda_m$ and $n = \ln N$. Take the natural logarithm to eqn 13 and obtain $r_t = G(r_{t-1}, n_{t-1})$ and $n_{t+1} = F(r_t, n_t)$ where the two functions G and F are defined as

$$\begin{aligned} G &= r_{t-1} - \gamma_q n_{t-1} + \sigma^2 \\ F &= r_t + (1 - \gamma)n_t \end{aligned} \quad (17)$$

Following the procedure in Bulmer (1994), the stability of the evolutionary equilibrium is given by the eigenvalues of the Jacobian matrix

$$\begin{bmatrix} \partial G/\partial r & \partial G/\partial n \\ \partial F/\partial r & \partial F/\partial n \end{bmatrix} \quad (18)$$

where the equilibrium is unstable when the absolute value of the dominant eigenvalue is larger than or equal to one. For the two-dimensional matrix eqn 18 the eigenvalues are

$$(T \pm \sqrt{T^2 - 4D})/2 \quad (19)$$

where D is the determinant and T the trace of that matrix. To determine D and T we have that

$D = (\partial G/\partial r)(\partial F/\partial n) - (\partial F/\partial r)(\partial G/\partial n)$, that $T = \partial G/\partial r + \partial F/\partial n$, and that $\partial G/\partial r = 1$, $\partial G/\partial n = -\gamma_q$, $\partial F/\partial r = 1$, and $\partial F/\partial n = 1 - \gamma$ for eqn 17 for all r and n including the equilibrium. Hence

$$\begin{aligned} D &= 1 - \gamma + \gamma_q \\ T &= 2 - \gamma \end{aligned} \quad (20)$$

and, thus, for eqn 13 on a logarithmic scale, the eigenvalues are

$$(2 - \gamma \pm \sqrt{\gamma^2 - 4\gamma_q})/2 \quad (21)$$

The eigenvalues are real when $\gamma^2 \geq 4\gamma_q$ and complex when $\gamma^2 < 4\gamma_q$. For the situation $\gamma^2 \geq 4\gamma_q$ with real eigenvalues the equilibrium is unstable when $\gamma > 4$ or when $\gamma_q \leq 2\gamma - 4$ and $2 \leq \gamma \leq 4$ while it is otherwise stable. When additive genetic variation is absent, i.e., when $\sigma^2 = 0$, the instability criterion $\gamma_q \leq 2\gamma - 4$ reduces to $\gamma \geq 2$ because $\gamma_q = 0$ at the limit where additive genetic variation is absent. This instability is generated by over-compensatory density regulation. This is in contrast to the instability criterion associated with the complex eigenvalues, where the instability arises from the evolutionary modulation of the growth rate. For the latter situation $\gamma^2 < 4\gamma_q$ and the absolute value of the eigenvalues is $\sqrt{1 - \gamma + \gamma_q}$ so that the evolutionary equilibrium is unstable when $\gamma \leq \gamma_q$. This latter instability criterion is impossible when additive genetic variation is absent, and this is because the absence of additive genetic variation imposes the constraint $\gamma_q = 0$ so that the inequality $\gamma \leq \gamma_q$ is false because γ is positive for natural situations.

To examine the conditions associated with the instability criterion $\gamma_q \geq \gamma$ recall that $\gamma_q = \gamma_i \psi \sigma^2$ and that $\gamma = \gamma_\alpha + \mu \gamma_i$. Therefore, the condition $\gamma_q \geq \gamma$ resembles

$$\gamma_\alpha/\gamma_i \leq \psi \sigma^2 - \mu \quad (22)$$

This implies that the equilibrium becomes more stable when the resource regulation (γ_α) or the average cost of interference (μ) are increased. Moreover, the equilibrium becomes more unstable when the density dependence in the level of interference (γ_i), the potential response to selection (σ^2), and the intra-population differentiation in interference regulation (ψ) are increased.

Let us now turn to the dynamics that follow from a perturbation of the evolutionary equilibrium of eqn 13. When additive genetic variation is present the long-term dynamics is given by the eigenvalues of the Jacobian matrix eqn 18. If these eigenvalues are real and their absolute values are smaller than one the perturbation will decline geometrically toward the equilibrium, while it will increase geometrically if the absolute value

of the dominant eigenvalue is larger than one. Hence, for eqn 13 the perturbation will increase geometrically when $\gamma > 4$ or when $\gamma_q \leq 2\gamma - 4$ and $2 \leq \gamma \leq 4$, while the dynamics is damped otherwise. Moreover, if the sign of the dominant eigenvalue is negative the long-term dynamics is oscillatory with a period of two generations, while it is monotonic if the value is positive. Hence, the long-term dynamics of eqn 13 is oscillatory when $\gamma > 2$ while it is monotonic for $\gamma < 2$. Numerical simulations though do show that the dynamics following from a perturbation of the equilibrium often is oscillatory even when $\gamma < 2$.

If instead the eigenvalues are complex they will induce a cycle with a highly variable period. To describe this cycle, let a complex eigenvalue be given by the polar co-ordinates in the Argand diagram, i.e., let x be its absolute value and θ its argument. Then the period of the cycle is $P = 2\pi/\theta$ while the amplitude is proportional both to the initial perturbation and to x . This amplitude is stable if $x = 1$, it declines geometrically if $x < 1$, and it increases geometrically if $x > 1$. That is, $x < 1$ corresponds to a damped cycle, $x = 1$ to a stable, and $x > 1$ to a repelling cycle. For eqn 13 we find that

$$\begin{aligned} x &= \sqrt{1 - \gamma + \gamma_q} \\ P &= 2\pi/\arctan \left| \frac{\sqrt{4\gamma_q - \gamma^2}}{2 - \gamma} \right| \end{aligned} \quad (23)$$

This implies that the cycle caused by the evolutionary modulation of the growth rate is damped when $\gamma < \gamma_q$, stable when $\gamma = \gamma_q$, and repelling when $\gamma > \gamma_q$. Also, when the cycle is stable the period has a minimum of 4 generations at $\gamma = 2$, while the period increases monotonically to infinity as γ goes toward either 0 or 4. For $\gamma < 0$ the system is not biologically defined, and at the limit of, and beyond, $\gamma = 4$ the eigenvalues are real with the dominant eigenvalue being smaller than -1 so that an initial perturbation increases geometrically and oscillatory without limits.

Let the period of the population cycle be the period between two neighbouring events where the population abundance crosses the equilibrium in the same direction. Numerical simulations show that the population period has a close resemblance to the period defined by eqn 23 when $\gamma_q \approx \gamma < 2$. When instead $\gamma_q \approx \gamma \geq 2$ and $\gamma_q \approx \gamma \rightarrow 4$ the population period declines monotonically from 4 to 2 generations, while the increasing period given by eqn 23 resembles a period that turns up in the amplitude of the population cycle. An example of the period in the amplitude is shown in Fig. 2, and the relation between the γ parameter and the population period is shown in Fig. 4 together with data

from forest insects.

B Parameter estimation

Given that the population cycle is stable the parameters for the population model eqn 13 are easily estimated from successive estimates of the population abundance: A stable cycle implies that $\gamma_q = \gamma$ and, thus, that eqns 13 and 13 reduces to

$$\begin{aligned}\lambda_{m,t} &= \lambda_{m,t-1} N_{t-1}^{-\gamma} e^{\sigma^2} \\ N_{t+1} &= N_t \lambda_{m,t} N_t^{-\gamma}\end{aligned}\quad (24)$$

To estimate the parameters γ and σ^2 , from the bottom equation of eqn 24, we have

$$\lambda_{m,t} = N_{t+1} N_t^{\gamma-1} \quad (25)$$

If we then insert $\lambda_{m,t-1}$ and $\lambda_{m,t}$ from eqn 25 into the top equation of eqn 24 we obtain

$$N_{t+1} N_t^{\gamma-1} = N_t N_{t-1}^{\gamma-1-\gamma} e^{\sigma^2} \quad (26)$$

This equation can be rearranged so that we can estimate γ and σ^2 by the linear regression

$$\ln N_{t+1} + \ln N_{t-1} = (2 - \gamma) \ln N_t + \sigma^2 \quad (27)$$

From $\lambda_m^{**} = \lambda_m^{**} N^{**-\gamma} e^{\sigma^2}$ [eqn 24] the abundance at the equilibrium is $N^{**} = e^{\sigma^2/\gamma}$, which combined with $N^{**} = N^{**} \lambda_m^{**} N^{**-\gamma}$ [eqn 24] gives us the estimate $\lambda_m^{**} = e^{\sigma^2}$.